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4 the environmental stress hypothesis

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15 <https://doi.org/10.6084/m9.figshare.26355463.v1> (Foest et al. 2025).

16 **Key words:** alternate bearing, climate gradients, climate marginality, intraspecific variation,  
17 masting, mast seeding, reproductive strategy

18 **Abstract**

19 Year-to-year variation in seed crop size (i.e. masting) varies strongly among populations of the  
20 same species. Understanding what causes this variation is vital, as masting affects the ability of  
21 tree species to regenerate and determines the population dynamics of a wide variety of animals.  
22 It is commonly thought that environmental stress is a key driver of masting variability. The  
23 environmental stress hypothesis posits that more marginal conditions increase the strength of  
24 masting. Using 437 time series from 19 tree species, we find that this hypothesis fails to fully  
25 explain how masting varies across marginality gradients. We expected higher interannual  
26 variation and less frequent masting events at species margins, but instead found that while mast  
27 years are indeed less frequent, the inter-annual variation was lower towards the margins. The  
28 observed patterns suggest that populations growing at the margins may invest more resources in  
29 low seed production years compared to their conspecifics, hedging their bets in these more  
30 challenging environments.

## 31 **Introduction**

32 For many long-lived plant species from diverse taxa and ecosystems, seed production is highly  
33 variable between years, and individuals within populations synchronise their reproductive effort  
34 (Pearse et al. 2016, Dale et al. 2021). The resulting population-level variability of seed  
35 production is known as masting (Kelly 1994, Pearse et al. 2016). A widely held hypothesis to  
36 explain masting variability is that plants growing in resource-poor, stressful or marginal  
37 environments show more pronounced masting (the ‘environmental stress’, or ‘productivity  
38 gradient’ hypothesis) (Kelly and Sork 2002, Satake and Bjørnstad 2008, Allen et al. 2012,  
39 Roland et al. 2014, Pearse et al. 2017, Wion et al. 2020, 2023). Here, we show that masting  
40 varies over environmental gradients, but that the environmental stress hypothesis does not  
41 explain this spatial variation.

42 We require a thorough understanding of what drives within-species variation in masting, given  
43 the important effects of masting on ecosystems. Masting increases reproductive efficiency  
44 through economies of scale that decrease seed predation rates and improve pollination efficiency  
45 (Kelly and Sork 2002, Pearse et al. 2016). Understanding these reproductive patterns in plants is  
46 increasingly important in the context of climate change, because plant reproduction is a critical  
47 component of forest dynamics following climate-driven disturbance and environmental change  
48 (Sharma et al. 2022). Climate-driven changes to mast seeding can severely limit the viability of  
49 seeds (Bogdziewicz et al. 2023). Moreover, seeds represent a foundational food source and  
50 masting affects animal populations throughout food webs (Clark et al. 2019).

51 The environmental stress hypothesis predicts that comparatively more stressful, lower  
52 productivity environments have less frequent large seed crops, resulting in higher inter-annual  
53 variability of seed production when compared to less limited populations (Kelly and Sork 2002,  
54 Satake and Bjørnstad 2008, Allen et al. 2012, Pearse et al. 2017). According to the hypothesis,

55 such environments limit how quickly plants can accumulate sufficient resources to produce a  
56 bumper crop. With increasingly challenging growing conditions, we thus expect lower seed  
57 production in years following high seed crops (i.e. stronger resource depletion), causing the  
58 temporal autocorrelation at lag one year to become more negative (Sork et al. 1993, Koenig and  
59 Knops 2000).

60       Until now, the environmental stress hypothesis has not been tested at a large scale. Testing  
61 is complicated by the fact that many factors can reduce plant productivity, and the stresses  
62 experienced by plants vary by species and location. This may help explain why the few studies  
63 that have examined the effects of different climate and productivity gradients on masting find  
64 contradicting responses, i.e. positive, negative or neutral responses (Table 1). We propose a way  
65 to find commonalities between species in their responses to stress gradients by using the concept  
66 of ‘climate marginality’.

67       The limit of a species’ range is a manifestation of its ecological niche (Sexton et al. 2009).  
68 If the species range edges are near equilibrium (i.e. the species can be found throughout most of  
69 its potential distribution), climate marginality theory predicts that populations at the climatic  
70 periphery more frequently experience the effect of biotic and abiotic limiting factors compared to  
71 populations living in core conditions, and that demographic performance would consequently be  
72 impaired (Sexton et al. 2009, Pironon et al. 2017, Kunstler et al. 2021). Changes in reproduction  
73 patterns towards climatically marginal sites could result from the limited availability of  
74 resources, or from higher demands on these resources for maintenance or overcoming increased  
75 competition (Gaston 2009, Roland et al. 2014), or because of more frequent extreme climate  
76 events that affect reproduction such as late spring frosts and summer droughts (Nussbaumer et al.  
77 2020, Journé et al. 2021, Willi and Van Buskirk 2022). All in all, we expect to observe changes

78 in masting behaviour across climate marginality gradients, with masting becoming stronger with  
79 increasing climate marginality.

80 Here, we present the first attempt to generalise, across 19 tree species, the effect of  
81 environmental stress on masting. While intraspecific variation in masting by populations is  
82 widely acknowledged (Herrera et al. 1998, Greene and Johnson 2004, Satake and Bjørnstad  
83 2008, Crone et al. 2011), it has remained largely unclear how much intraspecific variation of  
84 masting exists within species across space, and what drives this variability. Thus far, studies  
85 designed to assess the effect of environmental stress on intraspecific masting variability have  
86 been constrained in terms of spatial and species coverage. Now, due to a recent synthesis of data  
87 on reproductive behaviour in plants (Hackett-Pain et al. 2022), we are able to test the long-  
88 standing environmental stress hypothesis across species and extensive environmental gradients.

89 We expect to observe parabolic relationships between climate variables and masting  
90 metrics. Specifically, we expect (1) seed production to be more variable, (2) more negatively  
91 autocorrelated through time (indicative of a stronger resource limitation; Koenig et al. 2003), and  
92 (3) have less frequent large seed crops in populations growing in marginal climates. In contrast,  
93 range centres should be characterised by less variable, less autocorrelated and more frequent  
94 large seed crops.

## 95 **Materials and Methods**

96 We describe the variability of masting for populations from 19 tree species occurring in mid-to-  
97 high-latitudes using three masting metrics. Subsequently, we test if these metrics vary along  
98 climate marginality gradients in the ways predicted by the environmental stress hypothesis. For  
99 the six best-represented species in terms of data coverage, we also fit models to gain insight into  
100 the species-specific responses behind the overall, across species, pattern. All models were fitted  
101 in R v. 4.3.1 (R Core Team 2023).

## 102 **Intraspecific variation in reproductive patterns**

### 103 **Reproduction time series**

104 We obtained time series of annual reproduction from the open-access MASTREE+ database  
105 (Hackett-Pain et al. 2022). This database collates geo-referenced reproductive time series from  
106 perennial plant species across the world. We subset MASTREE+ to seed, fruit and cone  
107 production time series of species which had extensive temporal ( $\geq 10$  years of observation) and  
108 spatial ( $\geq 10$  sites) replication. From here on, seeds, fruit and cones will be referred to as seeds,  
109 for simplicity. *Quercus robur* and *Quercus petraea* time series were grouped together, as the  
110 often sympatric species have high hybridisation rates (Abadie et al. 2012), and synchronise their  
111 seed production (Bogdziewicz et al. 2017). We only included series representing stand and patch  
112 scale populations (excluding (super-)regional scale records) with high-precision spatial  
113 references, to accurately link reproductive behaviour and climate. Moreover, to ensure the  
114 masting metrics could be meaningfully calculated and compared, we excluded series which  
115 lacked a unit (i.e. ordinal data, index data) or which had synchrony-based units (i.e. ‘% of  
116 individuals reproducing’). Comparable collection methods were grouped together before  
117 conducting the analyses, and classified as either seed trap data, seed counts, or harvest records.  
118 This approach resulted in a dataset of 8,082 annual records, forming 437 time series of the  
119 reproductive behaviour of 19 plant species in 357 locations (Appendix S1: Table S1). On  
120 average, time series were 18.5 years long (median: 14, range: 10-62).

### 121 **Temporal reproductive patterns**

122 Three metrics capturing temporal reproductive patterns were calculated for each time series to  
123 describe masting variability. Firstly, we calculated the coefficient of variation  $CV_p$ , the standard  
124 deviation divided by the mean annual seed crop size (i.e. number of seeds). Temporal

125 autocorrelation was calculated with the 1-year lagged autocorrelation, AR(1), and obtained with  
126 the Acf function in R (Hyndman et al. 2022). Lastly, we calculated the proportion of high seed  
127 crop years,  $P_{sd}$ . High seed crop years (“mast years”) were defined as years when the  
128 standardised annual deviate of reproductive effort exceeded the absolute magnitude of the largest  
129 deviate below the mean, as proposed by LaMontagne and Boutin (2007). Additionally, we  
130 calculated the recently introduced  $kCV_p$ , a bounded alternative to  $CV_p$ , which was found to  
131 increase statistical power when compared with  $CV_p$  when testing for latitudinal patterns in  
132 masting (Lobry et al. 2023). This metric can be calculated from the  $CV_p$  by dividing the squared  
133  $CV_p$  by  $1 +$  the squared  $CV_p$  and subsequently taking the square root.

## 134 **Environmental gradients**

### 135 **Climate data and species’ ranges**

136 Climate data for each location were extracted from WorldClim v. 2.1 (Fick and Hijmans 2017).  
137 Temperature variables were adjusted (lapse rate:  $0.65\text{ }^{\circ}\text{C}/100\text{ m}$ ) when the elevation of  
138 MASTREE+ time series was known and deviated from the SRTM data used for WorldClim  
139 (Fick and Hijmans 2017). To assess the climate gradients covered by our sampling, we extracted  
140 climate data for species distributions. Species distributions for all but two species were obtained  
141 from EUFORGEN (EUFORGEN 2022) and the Atlas of United States Trees (Petty 2021). The  
142 range of *Fagus crenata* was digitised from Kobashi et al. (2006), and a digitised range of  
143 *Araucaria araucana* was provided by CONICET-UNS (M. Hadad, CONICET-UNS,  
144 unpublished data, 2023). For most species in the dataset, wide spatial and elevational gradients  
145 were sampled (Appendix S1: Fig. S1, Appendix S1: Table S1), capturing a large part of the  
146 climatic envelope (Appendix S1: Fig. S2, Appendix S1: Section S2).

147 **Linear models**

148 *Across-species patterns*

149 We tested for intraspecific effects of climate gradients on temporal reproductive patterns  
150 (i.e. CVp, AR1, Psd) using the glmmTMB package (v. 1.1.7; Brooks et al. 2017). To allow for  
151 comparisons between CVp and kCVp, we also substituted kCVp in the final CVp models.

152 While the inclusion of the species as a predictor captures between-species differences in  
153 the dependent variable, this does not account for potential between-species differences in the  
154 other predictors (Van de Pol and Wright 2009; Pearse et al. 2020). We therefore subject-centred  
155 all dependent and independent numeric variables by subtracting the species' (i.e. 'subject') mean  
156 of each respective variable, except for climate variables (N.B. models using this data are referred  
157 to as 'subject-centred models'). For climate variables, we subtracted the median value observed  
158 in the species' range, which allowed for explicit testing of our hypotheses, as the centred value  
159 captures distance from the core climate.

160 For each metric, we ran two models, namely (1) an annual average model, including mean  
161 annual temperature (MAT), annual precipitation (AP), the interaction between MAT and AP, as  
162 well the second order polynomials of MAT and AP as predictors, and (2) a seasonal model, using  
163 the temperatures and precipitation of the coldest (cq) and hottest (hq) quarter of the year, the  
164 temperature-precipitation interactions for both quarters (i.e.  $T_{cq} \times P_{cq}, T_{hq} \times P_{hq}$ ) and the  
165 second order polynomials of both temperature and precipitation variables. Since all sites are  
166 situated in mid-to-high latitudes, the hottest and coldest quarter correspond to summer and  
167 winter, respectively. We accounted for methodological differences and time series properties by  
168 including, as predictors (1) time series length, (2) a categorical classification of the measured  
169 reproductive variable (i.e. 'cone', 'fruit', 'seed'), and (3) the data collection method (i.e. 'harvest

170 record', 'seed trap', 'count-based method'). Since centring removes the average effect of species,  
171 no random factor was used to capture between-species differences. The AR(1) models showed a  
172 statistically significant latitudinal trend in the model residuals, and we therefore included  
173 subject-centred 'Latitude' as a predictor in these models. All numeric predictor variables were  
174 scaled using the root mean square to ensure model convergence. We allowed the dispersion of  
175 residuals to vary as a function of species, collection method and reproductive variable. Models  
176 were examined and validated with DHARMA (v. 0.4.6), by simulating model residuals and  
177 plotting these against observations and predictors to check for misfits, as well as testing for  
178 correct distribution (KS test), dispersion and outliers (Hartig and Lohse 2022). Additionally, we  
179 tested for collinearity with car (v. 3.1.2; Fox and Weisberg 2019). To assess the significance of  
180 the linear terms in these across-species models, we conducted likelihood ratio tests. Standard p-  
181 values reflect conditional hypotheses and are not suitable for evaluating the independent  
182 contributions of predictors (x) due to the inclusion of quadratic terms ( $I(x^2)$ ). Specifically, for  
183 each linear climate predictor, we compared a "full" model—i.e. the model described above but  
184 excluding the quadratic and interaction terms involving the focal predictor—to a "reduced"  
185 model in which the linear term itself was also removed (see Appendix S1: Table S2).

### 186 *Species-specific patterns*

187 Since mean trends may obscure ecologically relevant processes, we also fitted species-specific  
188 linear models with glmmTMB for CV<sub>p</sub>, AR(1) and Psd for six species with  $\geq 30$  observations.  
189 We used the same predictors as those used in the subject-centred models, but since only a single  
190 species was examined in each model, variables other than the climate variables were not species-  
191 centred. Variable scaling was not required for convergence. If there was no variation in  
192 reproductive variable or collection method for a species, the variable could not be included in the

193 model. The collection method predictor was removed from the *Picea glauca* models because it  
194 exceeded the collinearity threshold ( $GVI\overline{F^{2 \times df}} > \sqrt{10}$ ).

## 195 **Results**

### 196 **Intraspecific variation**

197 Reproductive patterns varied widely within species (Appendix S1: Table S1, Fig. 1, Fig. 2).  
198 Notably, for 14 out of 19 species, the observed intraspecific range of CVp values was greater  
199 than the interspecific variation in mean CVp (0.89-1.94, Appendix S1: Table S1, Fig. 1).  
200 Similarly, the intraspecific variation in AR(1) of 16 species exceeded the observed interspecific  
201 variation in mean AR(1) (-0.39-0.01, Appendix S1: Table S1, Fig. 1). The majority of time  
202 series, both within and across species, had negative AR(1) values. All intraspecific ranges in Psd  
203 exceeded the interspecific range in mean Psd (0.11-0.21, Appendix S1: Table S1, Fig. 1).

204 The species-specific metric space plots in Fig. 2 highlight how intraspecific variation in  
205 CVp and AR(1) in some species cover large portions of the across-species metric space (e.g. *A.*  
206 *incana*: 49.5%, *P. abies*: 32.9%). In contrast, some species have high intraspecific variability in  
207 only one metric (e.g. CVp of *P. engelmannii*, AR(1) of *P. sylvestris*), or relatively low  
208 intraspecific variability in both CVp and AR(1) (e.g. *A. amabilis*: 2.9%, *Q. douglasii*: 3.7%).

### 209 **Climate marginality**

#### 210 **Across-species patterns**

211 Large seeding years were less frequent towards more marginal climates, which is consistent with  
212 predictions of the environmental stress hypothesis. In the annual model, Psd responded to within-  
213 species change in MAT<sup>2</sup> (Fig. 3, climate effects: Table 2, full model: Appendix S1: Table S3).  
214 Sites with a climate closely matching the species' range median MAT displayed more frequent  
215 masting. It should be noted that while the concave relationship between Psd and AP best fitted

216 the observations, this relationship was not statistically significant. Squared winter temperatures  
217 predicted Psd in the seasonal model, with the lowest masting frequencies occurring at cold and  
218 warm sites. Thus, annual climate patterns appear to be more closely related to climatic conditions  
219 in the winter rather than the summer.

220 Yet, these changes in the frequency of large seeding years did not result in a higher CVp  
221 towards the margins, as predicted by the environmental stress hypothesis. CVp varied  
222 significantly with temperature, with the annual model predicting lower CVp in the warmest sites,  
223 and no effect of precipitation was observed (Fig. 3, climate effects: Table 2, full model:  
224 Appendix S1: Table S3). The seasonal model shows that the effect of winter temperature  
225 depends on winter precipitation. For populations growing in sites close to the species' median  
226 winter precipitation level, the CVp declined with increasing winter temperatures. Combined,  
227 these findings show that the relationships between CVp and marginality gradients are more  
228 complex than anticipated under the environmental stress hypothesis. The simulated kCVp model  
229 residuals showed deviations from normality (Kolmogorov-Smirnov; annual model:  $D = 0.08$ ,  $p <$   
230  $0.05$ ; seasonal model:  $D = 0.09$ ,  $p < 0.01$ ), but the model results are presented in Appendix S1:  
231 Fig. S3 and Appendix S1: Table S3 to allow for comparisons with the CVp models. The results  
232 from the annual average kCVp model imply that year-to-year variability in seed production  
233 decreases towards cold and warm margins, not just towards the warmer sites as in the CVp  
234 model. These findings directly contrast with the predictions under the environmental stress  
235 hypothesis. While the detected climate patterns differ from the annual average CVp model, the  
236 same methodological variables were found to have effects in both the annual average CVp and  
237 kCVp models (i.e. reproductive variable and data collection method). In the seasonal model, a  
238 similarly nuanced story emerges, and the observed relationships between climate and the

239 temporal variability of seed production also differed. A concave relationship was found between  
240 squared winter temperatures and  $kCV_p$ , rather than the significant interaction between winter  
241 temperature and precipitation detected in the  $CV_p$  model. Both models show that  $kCV_p$  is not  
242 simply a more powerful variant of  $CV_p$ . Instead, different relationships may be detected.

243       Lastly, how  $AR(1)$  varies along climate marginality gradients is not well explained by the  
244 predictions under the environmental stress hypothesis. Intraspecific variation in  $AR(1)$  varied  
245 with MAT in the annual climate model, where sites with higher mean annual temperatures had  
246 lower  $AR(1)$  values (Fig. 3, climate effects: Table 2, full model: Appendix S1: Table S3).  
247 Similarly to the annual model, the seasonal model indicates  $AR(1)$  was lower when sites  
248 experienced warmer temperatures in winter.  $AR(1)$  decreased towards both winter precipitation  
249 margins, i.e. towards dry and wet sites. The effect of the summer temperatures depended on  
250 precipitation; lower  $AR(1)$  occurred in drier and colder sites, as well as in warmer and wetter  
251 sites. Summer precipitation also showed a parabolic relationship with  $AR(1)$ , where  $AR(1)$  was  
252 lowest at the wettest and the driest sites. Thus, we only found limited evidence for concave  
253 relationships in the seasonal model.

254       Overall, we find that masting patterns vary across marginality gradients. However, this  
255 variation differs from predictions under the environmental stress hypothesis (Table 3). In  
256 addition to these diverging relationships, we also find that there is substantial, and potentially  
257 ecologically meaningful, variation around these predictions (see partial residuals, Fig. 3). This  
258 indicates that factors other than climate marginality or methodological differences play an  
259 important role in driving masting variability.

## 260 **Species-specific patterns**

261 There is always a risk that ecologically meaningful relationships go undetected if trends are only  
262 tested at a general level. Or, phrased differently, there is a chance that the hypothesis holds, but  
263 only for a subset of species. Consequently, we tested how masting metrics varied across annual  
264 climate marginality gradients for the six best-sampled species (Appendix S1: Table S4). Here,  
265 we reveal, where applicable, evidence of species-specific responses matching the predictions of  
266 the environmental stress hypothesis.

267 Evidence in line with our prediction that Psd would be lower in the climatic margins was  
268 found for *P. abies* and *P. sylvestris*. Specifically, we observed concave effects on Psd of MAT  
269 for *P. abies*, and of AP for *P. sylvestris*, although highest Psd values were found in sites with  
270 higher-than-median precipitation in this species.

271 The relationship between CVp and climate did not generally match the predicted convex  
272 pattern of highest variability in more marginal climates, with the exceptions of the effect of MAT  
273 on *F. sylvatica* and the effect of AP on *P. edulis* (although this later effect was not statistically  
274 significant ( $p = 0.06$ )).

275 For AR(1), the evidence in line with predictions is the concave relationship between AR(1)  
276 and annual precipitation in *Q. robur/petraea*, and between annual precipitation and temperature  
277 and AR(1) in *Picea abies*.

278 It is worth noting that these species-specific models have relatively small sample sizes  
279 (Appendix S1: Table S1), and therefore had lower power than the subject-centred models. Yet,  
280 these models reveal that species responses are diverse.

## 281 **Discussion**

282 The environmental stress hypothesis fails to explain intraspecific masting variation when  
283 evaluated across climate marginality gradients (Table 3). Under the environmental stress

284 hypothesis, we would predict higher interannual variation and, simultaneously, less frequent  
285 large seeding events at species' margins (Kelly and Sork 2002, Satake and Bjørnstad 2008, Allen  
286 et al. 2012, Roland et al. 2014, Pearse et al. 2017, Wion et al. 2020, 2023). Instead, we found that  
287 while high seeding years indeed became less frequent towards the margins, the interannual  
288 variation becomes lower instead of higher towards the warm margin (or towards warm and cold  
289 margins for the alternative kCVp metric). Combined, these patterns signal more constant seed  
290 production towards the margins, not stronger masting. The patterns in the temporal  
291 autocorrelation are also indicative of more constant seed production. When there are few large  
292 seeding events (which happens towards the warm margins), this average relationship is  
293 determined mostly by pairs of subsequent low reproduction years. If populations shift from zero  
294 or extremely low seed output in low seed crop years to increased seed production in some of  
295 these years, the AR(1) would decrease, but the overall seed crop would become more constant.  
296 Overall, our results provide solid support that masting changes across climate gradients, but not  
297 in the way expected under the environmental stress hypothesis.

298       Our work helps to develop a clearer picture of how reproduction varies across climate  
299 space. Various studies have found that the number of flowers, fruits and seeds produced at the  
300 edge of the geographic distribution range were comparable to those produced at the geographic  
301 centre (Abeli et al. 2014, Pironon et al. 2017). However, in masting populations, reproductive  
302 outcomes depend not only on the number of seeds produced, but also on the temporal variability  
303 of these seed crops (Bogdziewicz et al. 2023). The more constant reproduction observed towards  
304 climate margins has many potential consequences for ecosystem dynamics. For example, less  
305 frequent mast years of large-seeded species like beech and oak lead to declines in seed consumer  
306 populations (Touzot et al. 2020). In addition to the general, across species, trends in masting

307 variability across space, we find evidence that species-specific responses to climate gradients are  
308 diverse. This implies that the masting of some species is more responsive to climatic variation  
309 than others.

310 In this analysis, the spatial sampling was limited by the availability of long-term  
311 reproductive data. Future efforts which deliberately sample marginal populations over decades  
312 would be highly valuable, and would help to detect species-specific signals of stress on masting.  
313 However, such data are currently still limited. Regardless, our data were sampled across strong  
314 gradients, which, for the best-sampled species, cover the reported climatic envelope of the  
315 species (See Appendix S1: Section S2).

316 Our first large-scale assessment of within-species variation in masting shows that it can  
317 exceed the among-species variation. *Picea glauca*, for example, is a species that is typically  
318 associated with masting (i.e. high variability of seed production). Yet, we found that the CV<sub>p</sub> of  
319 populations of this species ranged from very high (3.34) to moderately low (0.93). This range is  
320 more than double the among-species range in mean CV<sub>p</sub> (0.89-1.94). Such high intraspecific  
321 variability provides further evidence of the pitfalls of classifying species as “masting” or “non-  
322 masting” (Kelly 2023), and illuminates the need for caution when working with species-level  
323 averages: such models implicitly assume that species means are meaningful, and that the  
324 intraspecific variation in the trait of interest is lower than the interspecific variation (Garamszegi  
325 2014, Westerband et al. 2021). Moreover, the striking extent of intraspecific variation in masting  
326 suggests that the downstream consequences of masting may be more spatially variable than  
327 assumed. Given the important effects of masting on wider ecosystems, it is crucial that we  
328 develop a thorough understanding of what causes masting to vary so much between populations.

329 Masting at species climate margins was characterised by less frequent large-seeding years  
330 and lower interannual variation in seed production. The observed patterns might be explained by  
331 larger reproductive investment in small seed production years towards species margins, as has  
332 been observed for European beech (Müller-Haubold et al. 2015). Theory predicts that the costs  
333 of masting, such as the risk of dying before producing bumper crops, increases with decreased  
334 adult survivorship, and that longer periods between large seeding years elevate these costs  
335 (Waller 1979, Journé et al. 2023). Therefore, lower survival rates at marginal populations  
336 together with lower frequencies of producing large seeding years may translate into higher  
337 reproductive investment in low seed production years, a form of bet hedging in these more  
338 challenging environments. Moreover, lower benefits to masting may also occur in these sites if  
339 seed predator pressures are lower (Alexander et al. 2007, Vaupel and Matthies 2012). Future  
340 work considering masting variability alongside factors determining the cost - balance equation of  
341 masting benefits is therefore required.

342 The observed variance of masting patterns around the effects of climate marginality  
343 suggests that factors beyond climate contribute to intraspecific masting variability. The  
344 importance of local environmental effects was demonstrated in *Pinus ponderosa*, where  
345 intraspecific variation in masting was associated with stand characteristics including stand  
346 density, tree size and age (Wion et al. 2023). Plant ontogeny was also found to be an important  
347 determinant of masting in some temperate forest trees (Pesendorfer et al. 2020). If these stand  
348 characteristics explain a large proportion of variability in masting, then the effects of climate  
349 change on masting and the availability of viable seeds at the climatic margins might be alleviated  
350 through appropriate management strategies (Bogdziewicz et al. 2023). A limited number of  
351 studies indicate that additional reproductive variability is explained by provenance (Mark 1965,

352 Caignard et al. 2021), but data from reciprocal transplant or common gardens established along  
353 climate gradients are still needed.

354 Due to the rapid accumulation of data on reproductive behaviours (Pearse et al. 2020,  
355 Hackett-Pain et al. 2022), we have been able to test the long-standing environmental stress  
356 hypothesis and unveil the remarkable within-species variation in masting. We find masting  
357 variation across climate space is inconsistent with the environmental stress hypothesis. High seed  
358 crop years are less frequent at the climatic periphery, but, importantly, the year-to-year  
359 variability of seed crops is lower towards the warm margin (or warm and cold margins for the  
360 alternative kCVp metric). As climate change may drive more populations to the climatic  
361 periphery, it is urgent that we establish if the spatial patterns observed here translate into  
362 temporal patterns. Recent work shows that this is not the case in European beech (Foest et al.  
363 2024). Moreover, more work is required to establish if changes in masting result in lower  
364 reproductive efficiency. Lower interannual variation at marginal populations may translate into  
365 less effective predator satiation, higher losses to seed predators, and lower reproductive  
366 efficiency (Zwolak et al. 2022). The large unexplained variation around the effects of climate not  
367 only sets important challenges, but also offers management potential (e.g. managing stand age  
368 and structure). If the source of this variability results from stand characteristics, we may be able  
369 to develop appropriate management strategies to optimise reproductive efficiency.

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### 390 **Author Contributions**

391 JJF, AH-P and TC conceptualised the study, and JJF led the development of the methodology,  
392 with support from AH-P, IP and TC. Investigation, formal analysis and data visualisation were  
393 performed by JJF. JJF wrote the original draft, and all authors contributed critical feedback on  
394 the manuscript. AH-P provided supervision.

### 395 **Conflict of interest statement**

396 The authors declare to have no competing interests.

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555 **Table section**

556 *Table 1: Relationships between environmental gradients and metrics capturing temporal*  
 557 *reproductive patterns per tree species reported in the literature.*

| Metric  | Gradient           | Direction of effect                                      |   |   | Species  |
|---------|--------------------|--|---|---|--|
|         |                    | Decrease   | Neutral   | Increase  |  |
| CVp     | Elevation          | Masaki <i>et al.</i> (2020); Roland <i>et al.</i> (2014) | Lázaro <i>et al.</i> (2006); Masaki <i>et al.</i> (2020); Mencuccini <i>et al.</i> (1995); Mooney <i>et al.</i> (2011)    | Buechling <i>et al.</i> (2016); Kelly <i>et al.</i> (2001); Kelly <i>et al.</i> (2008); Masaki <i>et al.</i> (2020); Sullivan and Kelly (2000); Webb and Kelly (1993) | <i>Fagus crenata</i> , <i>Swida controversa</i> ; <i>Picea glauca</i><br><br><i>Buxus balearica</i> ; <i>Quercus serrata</i> , <i>Castanea crenata</i> , <i>Prunus grayana</i> ; <i>Picea abies</i> ; <i>Pinus ponderosa</i> |
|         |                    |  |   |   |  |
|         | Water availability | Espelta <i>et al.</i> (2008); Wion <i>et al.</i> (2020)  | Abies, <i>Picea</i> , <i>Pinus</i> , <i>Tsuga</i><br>Global scale<br><i>Chionochloa spp.</i> ; <i>Nothofagus solandri</i> |   |  |
|         |                    |  |   | Water availability  | Espelta <i>et al.</i> (2008); Wion <i>et al.</i> (2020)  |
| CVi & S | Water availability | Wion <i>et al.</i> (2023)                                | <i>Pinus ponderosa</i>  |   |  |

|                |                    |                                 |   |
|----------------|--------------------|---------------------------------|---|
| AR(1)          | Water availability | Barringer <i>et al.</i> (2013)  | <i>Quercus lobata, Quercus agrifolia</i>                  |
| Periodicity    | Elevation          | Allen <i>et al.</i> (2012)      | <i>Nothofagus solandri</i>                                |
| Mast frequency | Elevation          | Mencuccini <i>et al.</i> (1995) | <i>Picea abies</i>  |
|                | Nitrogen           |                                 | Tanentzap <i>et al.</i> (2012)<br><i>Chionochloa spp.</i> |

*Reference details in Appendix S1: Section S1. CV<sub>p</sub>: coefficient of variation of population-level seed crop size; CV<sub>i</sub>: coefficient of variation of tree-level seed crop size; S = between-tree synchrony; AR(1): 1-year lagged autocorrelation.*

559 *Table 2: Relationships between climate variables and masting metrics CVp, AR(1) and Psd*  
 560 *found using subject-centred annual average and seasonal models.*

| Term              | CVp   |              |             | AR(1)   |              |                 | Psd   |              |                 |
|-------------------|---|--------------|-------------|---|--------------|-----------------|---|--------------|-----------------|
|                   | Estimate  | z            | P-value     | Estimate  | z            | P-value         | Estimate  | z            | P-value         |
| <i>Annual</i>     |   |              |             |   |              |                 |   |              |                 |
| AP                | $9.73 \times 10^{-3}$<br>( $3.07 \times 10^{-2}$ )                                    | 0.32         | 0.75        | $1.49 \times 10^{-3}$<br>( $1.29 \times 10^{-2}$ )                                    | 0.12         | 0.91            | <b><math>9.67 \times 10^{-3}</math></b><br><b>(<math>4.51 \times 10^{-3}</math>)</b>  | <b>2.14</b>  | <b>0.03</b>     |
| AP <sup>2</sup>   | $-1.40 \times 10^{-3}$<br>( $1.38 \times 10^{-2}$ )                                   | -0.10        | 0.92        | $-2.86 \times 10^{-3}$<br>( $5.32 \times 10^{-3}$ )                                   | -0.54        | 0.59            | $-3.75 \times 10^{-3}$<br>( $2.03 \times 10^{-3}$ )                                   | -1.84        | 0.07            |
| MAT               | <b><math>-6.21 \times 10^{-2}</math></b><br><b>(<math>2.69 \times 10^{-2}</math>)</b> | <b>-2.31</b> | <b>0.02</b> | <b><math>-3.45 \times 10^{-2}</math></b><br><b>(<math>1.10 \times 10^{-2}</math>)</b> | <b>-3.13</b> | <b>&lt;0.01</b> | $-1.31 \times 10^{-3}$<br>( $4.08 \times 10^{-3}$ )                                   | -0.32        | 0.75            |
| MAT<br>× AP       | $1.43 \times 10^{-2}$<br>( $1.95 \times 10^{-2}$ )                                    | 0.73         | 0.47        | $-1.51 \times 10^{-2}$<br>( $9.75 \times 10^{-3}$ )                                   | -1.55        | 0.12            | $2.37 \times 10^{-3}$<br>( $3.55 \times 10^{-3}$ )                                    | 0.67         | 0.50            |
| MAT <sup>2</sup>  | $-8.66 \times 10^{-3}$<br>( $8.77 \times 10^{-3}$ )                                   | -0.99        | 0.32        | $9.95 \times 10^{-3}$<br>( $5.95 \times 10^{-3}$ )                                    | 1.67         | 0.09            | <b><math>-4.23 \times 10^{-3}</math></b><br><b>(<math>2.06 \times 10^{-3}</math>)</b> | <b>-2.05</b> | <b>0.04</b>     |
| <i>Seasonal</i>   |   |              |             |   |              |                 |   |              |                 |
| hq P              | $-4.08 \times 10^{-3}$<br>( $2.77 \times 10^{-2}$ )                                   | -0.15        | 0.88        | $2.29 \times 10^{-3}$<br>( $1.09 \times 10^{-2}$ )                                    | 0.21         | 0.83            | $-1.74 \times 10^{-3}$<br>( $3.81 \times 10^{-3}$ )                                   | -0.46        | 0.65            |
| hq P <sup>2</sup> | $6.14 \times 10^{-3}$<br>( $1.22 \times 10^{-2}$ )                                    | 0.50         | 0.61        | <b><math>-1.14 \times 10^{-2}</math></b><br><b>(<math>3.98 \times 10^{-3}</math>)</b> | <b>-2.87</b> | <b>&lt;0.01</b> | $1.35 \times 10^{-3}$<br>( $1.60 \times 10^{-3}$ )                                    | 0.84         | 0.40            |
| hq T              | $5.15 \times 10^{-3}$<br>( $2.44 \times 10^{-2}$ )                                    | 0.21         | 0.83        | $2.32 \times 10^{-4}$<br>( $8.66 \times 10^{-3}$ )                                    | 0.03         | 0.98            | $-1.07 \times 10^{-3}$<br>( $3.21 \times 10^{-3}$ )                                   | -0.33        | 0.74            |
| hq T ×<br>hq P    | $-4.49 \times 10^{-3}$<br>( $1.83 \times 10^{-2}$ )                                   | -0.25        | 0.81        | <b><math>-2.18 \times 10^{-2}</math></b><br><b>(<math>7.17 \times 10^{-3}</math>)</b> | <b>-3.05</b> | <b>&lt;0.01</b> | $-2.91 \times 10^{-3}$<br>( $2.70 \times 10^{-3}$ )                                   | -1.08        | 0.28            |
| hq T <sup>2</sup> | $-5.03 \times 10^{-3}$<br>( $1.33 \times 10^{-2}$ )                                   | -0.38        | 0.70        | $-1.79 \times 10^{-3}$<br>( $5.07 \times 10^{-3}$ )                                   | -0.35        | 0.72            | $1.07 \times 10^{-3}$<br>( $1.91 \times 10^{-3}$ )                                    | 0.56         | 0.58            |
| cq P              | $1.96 \times 10^{-2}$<br>( $3.15 \times 10^{-2}$ )                                    | 0.62         | 0.54        | $1.78 \times 10^{-2}$<br>( $1.25 \times 10^{-2}$ )                                    | 1.43         | 0.15            | $6.74 \times 10^{-3}$<br>( $4.87 \times 10^{-3}$ )                                    | 1.39         | 0.17            |
| cq P <sup>2</sup> | $-1.36 \times 10^{-2}$<br>( $1.01 \times 10^{-2}$ )                                   | -1.35        | 0.18        | <b><math>-8.32 \times 10^{-3}</math></b><br><b>(<math>3.79 \times 10^{-3}</math>)</b> | <b>-2.20</b> | <b>0.03</b>     | $-2.23 \times 10^{-3}$<br>( $1.53 \times 10^{-3}$ )                                   | -1.46        | 0.14            |
| cq T              | <b><math>-1.02 \times 10^{-1}</math></b><br><b>(<math>4.45 \times 10^{-2}</math>)</b> | <b>-2.30</b> | <b>0.02</b> | <b><math>-5.96 \times 10^{-2}</math></b><br><b>(<math>1.92 \times 10^{-2}</math>)</b> | <b>-3.10</b> | <b>&lt;0.01</b> | $8.26 \times 10^{-3}$<br>( $5.98 \times 10^{-3}$ )                                    | 1.38         | 0.17            |
| cq T ×<br>cq P    | <b><math>4.46 \times 10^{-2}</math></b><br><b>(<math>2.11 \times 10^{-2}</math>)</b>  | <b>2.11</b>  | <b>0.03</b> | $8.31 \times 10^{-3}$<br>( $1.40 \times 10^{-2}$ )                                    | 0.59         | 0.55            | $3.42 \times 10^{-3}$<br>( $4.53 \times 10^{-3}$ )                                    | 0.75         | 0.45            |
| cq T <sup>2</sup> | $-4.68 \times 10^{-3}$<br>( $1.45 \times 10^{-2}$ )                                   | -0.32        | 0.75        | $7.67 \times 10^{-3}$<br>( $8.76 \times 10^{-3}$ )                                    | 0.88         | 0.38            | <b><math>-8.13 \times 10^{-3}</math></b><br><b>(<math>2.53 \times 10^{-3}</math>)</b> | <b>-3.21</b> | <b>&lt;0.01</b> |

*Standard errors within parentheses. P-values were calculated with Wald tests via the summary() function. For linear predictors, these p-values reflect conditional hypotheses and should not be interpreted as evidence of independent contributions due to the inclusion of quadratic terms as I(x^2). Significant effects are in bold, and underlined values indicate that the independent linear predictor was also significant in likelihood ratio tests comparing a model with the predictor (excluding interactions and quadratic terms) to one without it, to test the independent contributions. These likelihood ratio results can be found in Appendix S1:*

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*Table S2. Abbreviations = CVp: coefficient of variation of population-level seed crop size; AR(1): 1-year lagged autocorrelation; Psd: the proportion of large seed production years; AP: annual precipitation; MAT: mean annual temperature; hq: hottest quarter; cq: coldest quarter; T: temperature; P: precipitation.*

562 *Table 3: Summary of main findings, showing if the relationship between climate and masting*  
 563 *metrics matched predictions under the environmental stress hypothesis.*

| Environmental stress hypothesis |   |                  |           |   |
|---------------------------------|---|------------------|-----------|---|
| Metric                          | Prediction  | Climate variable | Supported | Results   |
| Psd                             | Less frequent masting in more marginal climates         | MAT              | Yes       | Less frequent masting in cool & warm sites                    |
|                                 |   | AP               | Partially | Less frequent masting in dry & wet sites, but not significant |
| AR(1)                           | More negative autocorrelation in more marginal climates | MAT              | No        | More negative autocorrelation in warm sites                   |
|                                 |   | AP               | No        | No significant effects  |
| CVp                             | Higher variability in more marginal climates            | MAT              | No        | Lower variability in warm sites                               |
|                                 |   | AP               | No        | No significant effects  |
| kCVp                            | Higher variability in more marginal climates            | MAT              | No        | Lower variability in cold & warm sites                        |
|                                 |   | AP               | No        | No significant effects  |

*Psd: the proportion of large seed production years; AR(1): 1-year lagged autocorrelation;*

*CVp: coefficient of variation of population-level seed crop size; AP: annual precipitation;*

*MAT: mean annual temperature.*

564

565 **Figure caption section**

566 *Figure 1: Intraspecific variation in masting metrics often exceeds interspecific variation. The*  
567 *shaded grey rectangle indicates the interspecific range in metric values. The colours of the violin*  
568 *plots indicate plant families, with blue-greens reserved for gymnosperms and orange-reds for*  
569 *angiosperms. CVp: coefficient of variation of population-level seed crop size; AR(1): 1-year*  
570 *lagged autocorrelation; Psd: the proportion of large seed production years. See Fig. 2 for genus*  
571 *names.*

572 *Figure 2: Within-species variation in masting metrics covers a large part of across-species*  
573 *masting metric space for some, but not all tree species. The top-left plot shows how three metrics*  
574 *co-vary (CVp: coefficient of variation of population-level seed crop size; AR(1): 1-year lagged*  
575 *autocorrelation; Psd: the proportion of large seed production years). Dashed-line convex hull:*  
576 *variation of CVp and AR(1) across all analysed species. Solid-line convex hull: the CVp and*  
577 *AR(1) metric space of all 378 species (1,018 time series) in MASTREE+ with time series*  
578 *matching subsetting criteria, irrespective of whether they had adequate spatial replication for*  
579 *inclusion in our analyses. Species-specific metric space is shown in surrounding plots.*  
580 *Percentages in the top-right corner of plots show the ratio of the solid to dashed convex hull*  
581 *area. Outlined time series in the top-left plot are plotted in the ‘Subsample’ plots (3 random time*  
582 *series [1960-2020] per region; regions marked with 1-3 and A-C) to illustrate example time*  
583 *series. Example time series are scaled by their maximum value.*

584 *Figure 3: Climate marginality explains variation in masting, but patterns diverge from*  
585 *predictions under the environmental stress hypothesis. The three columns on the left show*  
586 *response surfaces for the subject-centred models of masting metrics CVp, AR(1) and Psd (CVp:*  
587 *coefficient of variation of population-level seed crop size; AR(1): 1-year lagged autocorrelation;*

588 *Psd: the proportion of large seed production years) at baseline levels of the centred and scaled*  
589 *predictor variables. The annual average model is found in the first column, and the results of the*  
590 *seasonal model (hq: hottest quarter; cq coldest quarter; T: temperature, P: precipitation) are*  
591 *found in columns 2-3. The boundaries of the response surfaces in columns 1-3 were determined*  
592 *by a convex hull of climate data of the time series (open circles). The right-most column shows*  
593 *an alternative representation of the annual model found in the first column. Specifically, it shows*  
594 *the marginal effects (and associated 95% confidence intervals, shaded) for mean annual*  
595 *temperature (MAT) at different levels of annual precipitation (AP; blue colours) in the annual*  
596 *CVp, AR(1) and Psd models, alongside partial residuals. Spp.: species.*